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## Graphical Abstract

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## Highlights

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- Preferential sampling can be modeled by Poisson or conditional Poisson sampling.
- We document the impact of preferential sampling on population mean estimation.
- Preferential sampling can lead to biased mean estimation, if not accounted for.
- The bias increases with the covariance between sample membership and the variable.
- The bias decreases with increasing sampling effort (expected sampling fraction).

## On the impact of preferential sampling on ecological status and trend assessment

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#### Abstract

Assessments of the status and trends of abiotic and biotic indicators are two central objectives in many ecological studies and monitoring programs. Given the impracticality of making measurements or observations at every point in geographic space, even within a limited domain, consideration of spatial sampling is crucial to ensure the reliability of statistical inference regarding such status or temporal trends.

The sampling units in geographic space (e.g., sites, plots, quadrats) for field observations are often selected with a preference for those expected to be species-rich or those with the highest abundances or occupancy probabilities. This sampling approach, called *preferential sampling*, can be based on probability sampling theory, but in practice, it is usually a form of nonprobability sampling.

Introducing a selection force that disproportionately includes units in the sample based on the expected values of the variables of interest can lead to (severely) biased inferences. This is because inclusion probabilities — referred to here as *propensities* for units to be part of the sample — cannot be accounted for in statistical estimators when they are unknown to the sampler.

In this article, we model sampling processes (considered without replacement) for a finite spatial population of sampling units using probability sampling designs. We consider four designs: Bernoulli sampling, Poisson sampling, simple random sampling, and conditional Poisson sampling. We document the bias introduced by preferential sampling in the estimation of a mean, whether for a status assessment (e.g., mean species richness) or a trend assessment (e.g., trend in mean abundance). For this purpose, we use Monte Carlo simulations and an analytical expression for the bias of the sample mean.

This analytical expression shows that the bias of the sample mean (1) increases with increasing covariance between the propensities and the values of the variable of interest and (2) decreases with increasing sampling effort (sampling fraction or expected sampling fraction). This fundamental statistical result is neither widely known nor appreciated by most ecologists, even though it has the potential to ruin status or trend assessments and to lead to erroneous conclusions.

The findings on preferential sampling in ecology presented in this article are reviewed from a methodological perspective, mainly for an audience of quantitative ecologists, wildlife statisticians, and biometricians involved in the design or implementation of ecological studies and monitoring programs. To facilitate future exchange among researchers on this topic by clarifying the concepts, in the discussion we also examine the terminology found in the literature for the notions related to preferential sampling.

#### Keywords:

biased site selection, spatial sampling processes, propensities, range shift, abundance gradient, conditional Poisson sampling

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#### 1 1. Introduction

Although it has been abandoned in modern physics, the separation of space and time proves convenient and operational at macroscopic scales relevant to ecology. This allows us to consider all ecological phenomena as occurring in space and time, whether these dimensions are considered jointly

<sup>5</sup> or separately in their description, analysis and modeling.

 $_{6}$  In practice, the measurements or observations made, whether on abiotic or biotic variables, concern

 $_{7}$  a given spatial domain (which we denote  $\mathcal{D}$ ) and which take place over a given period (which we denote

 $_{*}$   $\mathcal{T}$ ). Even if  $\mathcal{D}$  and  $\mathcal{T}$  are not large, it is generally impossible to make measurements or observations

 $_{9}$  at every point in  $\mathcal{D}$  or every time in  $\mathcal{T}$ . Thus, a fundamental sampling problem immediately arises.

In the following, we assume that  $\mathcal{D}$  is a two-dimensional spatial domain lying in Euclidean space. Although not mandatory,  $\mathcal{D}$  is often partitioned into a finite population of areal units (see, e.g., Aubry and Francesiaz, 2022, Fig. 1). The same is true for the time domain, which can be discretized into a finite population of time units (e.g., potential daily counting sessions for bird counting).

Regardless of whether the population of spatiotemporal units is finite or considered as infinite, data collection is performed in two steps: (i) by selecting a subset of units from  $\mathcal{D} \times \mathcal{T}$  to form a sample *s* and (ii) by making measurements or observations on the units of *s*.

In the technical mathematical sense, an *error* is the deviation between a state (or a summary 17 of a state) defined on  $\mathcal{D} \times \mathcal{T}$  and its estimate or prediction computed from the sample s at hand. 18 The first step mentioned above introduces a sampling error because only a portion of the sampled 19 spatiotemporal population is considered when collecting the data. The second step generates an error 20 that can be treated as a measurement or observation error, for instance, when counting individuals 21 (e.g., Aubry et al., 2012) or when assessing the size of vast groups of wild animals (e.g., Vallecillo et al., 22 2021). When individuals are counted, this error can also be treated as a sampling error — where the 23 set of individuals counted is a sample of all individuals present — due to *imperfect detection* (e.g., 24 White, 2005; Kellner and Swihart, 2014; Perret et al., 2023; Johnston et al., 2023, Sec. 4.2). In this 25 article, we will focus only on the first source of error, concentrating on the sampling of  $\mathcal{D}$ . Since we are 26 interested only in spatial sampling, we do not discuss methods for counting individuals or problems of 27 imperfect detection or of imperfect ability to identify and to count them when they are detected (see. 28 e.g., the references cited by Vallecillo et al., 2021 and Aubry et al., 2023). 29

In a general sense, we call the sampling process the way by which a sample s is selected from 30 a statistical population, regardless of the method used. We reduce the set of sampling processes 31 by assuming here that they are without replacement. We assume that the sampling process under 32 consideration can be replicated — at least in principle — so that distinct samples (see Hedavat and 33 Sinha, 1991, p. 2) can be obtained on the basis of selection probabilities 0 < p(s) < 1. The individuals 34 or groups of individuals who perform the sampling may be collectively referred to as the sampler. 35 The sampling processes considered in this article are therefore human-based and, as such, are not 36 necessarily easy to model in detail (but see ter Steege et al., 2011; Fernández and Nakamura, 2015). 37

Statistical inference from a sample can be used to estimate parameters of an actual finite statistical 38 population or parameters of a hypothetical superpopulation describing the stochastic process under 39 study (see, e.g., Aubry and Francesiaz, 2022, Sec. 2.3). Sampling is said to be preferential when the 40 sampling process is related to a superpopulation (e.g., Diggle and Ribeiro, 2007, Sec. 4.4.2; Diggle 41 et al., 2010; Watson et al., 2019; Gray and Evangelou, 2023) or to the finite population at hand in such 42 a way that units are more likely to be included in the sample if they have the highest values for the 43 indicator of interest. For example, in ecology, preferential sampling may refer to preferential inclusion 44 in the sample of the most species-rich units, the units with the highest abundances or occupancy 45 probabilities. Conversely, we use the expression *antipreferential sampling* to denote situations where 46 units with the lowest values are more likely to be included in the sample. In all the other cases where 47 the probability of inclusion in the sample is independent of the characteristics of the unit with respect 48 to the variables of interest, sampling is said to be nonpreferential. 49

With the previous definition in mind, preferential sampling should be distinguished from (i) conve-50 nience sampling (sometimes called accessibility sampling; see, e.g., Young and Young, 1998, p. 93 or 51 Barnett, 2002, p. 17) and (ii) purposive sampling (also known as judgment sampling). In the first case, 52 units are selected for inclusion in the sample for convenience (e.g., ease of access and/or safety for the 53 observers). In the second case, units judged to be typical or appropriate for the survey are selected 54 from the statistical population under study, for example, when "ecologists actively seek the event of 55 interest, such as an active breeding nest of a bird, or a specific plant species" (Edwards et al., 2006). 56 The previous two sampling methods correspond to two forms of nonprobability sampling. Preferential 57 sampling can be either a nonprobability sampling method or be based on a random selection mecha-58 nism. We distinguish between the two cases (i.e., nonprobability vs. probability preferential sampling) 59 because the statistical implications are quite different. Indeed, in the first case, statistical inference 60

- $_{61}$  must be based on a model which can be implicit and its key assumptions, while in the second
- <sup>62</sup> case, the properties of the estimator used can be based solely on the random selection mechanism (see,

e.g., Aubry and Francesiaz, 2022).

- As a concrete situation, we assume in this article that sampling is performed in a nonprobability 64 setting. We consider two objectives that are central to many ecological studies or monitoring programs: 65 assessment of (i) status (i.e., the state observed at a given point in time) or (ii) trend (i.e., a smooth 66 pattern of state variation over time) (see, e.g., Gitzen et al., 2012). We do not discuss the scientific 67 relevance of either of these depending on the context in which they occur (see, e.g., Vos et al., 2000; 68 Yoccoz et al., 2001, 2003; Nichols and Williams, 2006). We merely consider status and trend as 69 objects of statistical inference, and our aim is to examine the impact of (nonprobability) preferential 70 (spatial) sampling on their assessment, mainly in terms of estimation bias. To this end (i) we formalize 71 preferential sampling in the context of finite population sampling; (ii) we briefly consider the one-stage 72 probability sampling designs that can be used to model basic without-replacement sampling processes; 73 (iii) using appropriate sampling process models, we examine the case of status and trend estimation 74 under preferential sampling through Monte Carlo simulation studies; (iv) we introduce to ecology a 75
- <sup>76</sup> fundamental formula that enables us to understand the very nature of the mean estimation bias that

 $\pi$  can be caused by preferential sampling. Finally, we draw lessons from the results presented, particularly

<sup>78</sup> with respect to trend assessment. Fig. 1 is intended to provide the reader with an overview of the

<sup>79</sup> article organization and of the topics covered before the discussion and perspectives sections.



Figure 1: Overview of the article organization before the discussion and perspectives sections. We first formalize preferential sampling in the context of finite population sampling. Next, we consider the probability sampling designs that can be used to model basic nonpreferential/preferential sampling processes. With these models, we then use Monte Carlo simulations to study the impact of preferential sampling, first for an example of status assessment, second for an example of trend assessment. The type of results to be examined is indicated to the right, either in a fixed population context or in a superpopulation context.

#### 2. Formalization 80

From a population of spatial units (e.g., sites, plots, quadrats)  $\mathcal{U}$  of size N, spatial sampling consists 81 of selecting a subset  $s \subseteq \mathcal{U}$  of size  $n_s \leq N$ , called a sample. The sampled spatial population should 82 not be confused with a biological population, i.e., here a set of organisms present in the domain  $\mathcal{D}$ . 83

The size of the spatial population N is not necessarily known to the sampler. In addition, the 84 sample size  $n_s$  may be variable or fixed. It is variable, for example, when the sample is drawn by a 85 group of people who do not necessarily consult each other and who each add one or more units to 86 the sample. The sample size can be fixed if the sample selection is more centralized, for example, to 87 predetermine the sampling effort to be allocated. 88

In the following, we consider a univariate situation, i.e., there is only one variable of interest y, 89 which takes fixed values for the units  $i \in \mathcal{U}$ . We denote  $\mathbf{y} = (y_1, y_2, \dots, y_N)$  as the vector of y-values for 90  $i \in \mathcal{U}$ . By extension, this vector can be called the population (implicitly of y-values) associated with the 91 population of spatial sampling units  $\mathcal{U}$ . The **y**-vector results from a bioecological stochastic process, 92 which may be modeled if appropriate. The observed **y**-vector is then assumed to be a realization of a 93 random vector (i.e., a superpopulation).

94

#### 2.1. Inclusion probabilities 95

In any sampling process, the probability  $0 < \pi_i \leq 1$  that a unit  $i \in \mathcal{U}$  is part of s: (i) is a probability 96 of inclusion known to the sampler, as is the case in finite population sampling theory (e.g., Särndal 97 et al., 1992; Tillé, 2020); or (ii) expresses the greater or lesser propensity of a unit to be part of the 98 sample, which is unknown to the sampler. In the latter case, the probabilities  $\pi_i$   $(i \in \mathcal{U})$  are propensity 99 scores, here simply called *propensities*. We denote  $\boldsymbol{\pi} = (\pi_1, \pi_2, \dots, \pi_N)$  as the vector of probabilities 100 for  $i \in \mathcal{U}$ . Let us note that the case  $\pi_i = 0$  corresponds to the exclusion of unit i from the sampled 101 population, a trivial case not of interest here. 102

Whether we are dealing with inclusion probabilities in the strict sense or with propensities, the  $\pi$ -103 vector results from the sampler's view of the variation of the variable of interest across geographic space, 104 without actually knowing it precisely. In particular, the probabilities considered in this article are not 105 modified when y-values are observed; therefore, the sampling process can be considered nonadaptive. 106 Thus, in this article, we assume that the samples are generated by a sampling process governed by 107 probabilities that are fixed in advance, whether consciously or not. 108

Where appropriate, the probabilities  $\pi_i$  can be defined as being proportional to a variable z ( $z_i > 0$ 109 for all  $i \in \mathcal{U}$ , which assigns more or less importance to the units — this type of variable is known as 110 a size variable — and is assumed to be correlated to some level with the variable of interest y. 111

#### 2.2. Preferential sampling 112

The sampling process is said to be preferential when there is a monotonically increasing relation 113 between the y-values and the inclusion probabilities, whether the latter are known or unknown. If the 114 relation decreases, we have antipreferential sampling. 115

We do not assume that the relation is linear, as this assumption is too restrictive to be realistic. 116 Indeed, a linear relation assumes that the propensity increases at a constant rate as a function of the y-variable, whereas in reality, it may increase faster at higher y-values, for example. Moreover, the 118 relation in question is not necessarily analytical; it can be only statistical — corresponding to the 119 conditional expectation — if the propensity varies for each y-value. 120

A necessary condition for sampling to be preferential (or antipreferential) is that the probabilities 121  $\pi_i$   $(i \in \mathcal{U})$  differ significantly between units or at least between groups of units. Sampling is inherently 122 nonpreferential if all units have, at least approximately, the same probability of being included in the 123 sample. 124

#### 2.3. Notation 125

In any sampling problem, we can distinguish at least two sources of stochasticity (see, e.g., Aubry 126 and Francesiaz, 2022): (i) one that relates only to the sampling process p(s) of the finite population 127 of units, resulting in the  $\pi$ -vector and (ii) the other related to the stochastic process  $\xi$  at the origin 128 of the y-vector (superpopulation). To avoid any ambiguity, operators (expectation, variance, etc.) are 129 denoted by using p as a subscript for the first source and  $\xi$  for the second. For a parameter  $\omega$ , the 130 sampling distribution of an estimator  $\hat{\omega}$ , i.e., the distribution of all the values that  $\hat{\omega}$  can take for 131 all the samples that can be generated, is called the *p*-distribution. The mean of the *p*-distribution is 132 called the *p*-expectation of  $\hat{\omega}$ , its variance is the *p*-variance (i.e., the sampling variance), and so on. 133 We define the p-bias as  $B_p(\hat{\omega}) = E_p(\hat{\omega}) - \omega$  and the relative p-bias as  $B_p(\hat{\omega})/\omega$ . For the stochastic 134 process  $\xi$ , we similarly denote the  $\xi$ -expectation,  $\xi$ -variance, and so on. When we consider both 135

<sup>136</sup> sources of stochasticity at the same time, we associate the two subscripts, and we write, for example,

 $_{137}$   $\xi p$ -expectation. Let us recall that in this article, in the context of double stochasticity,  $\pi$  is fixed, while

<sup>138</sup> y is random. Other notations used in this article are detailed in Appendix A.

#### 139 2.4. Parameters and estimators

Regardless of whether the spatial population size N is known to the sampler, for a variable of interest y (discrete, including binary, or continuous), we assume that the parameter to be estimated is the population mean  $\overline{y}_{II}$ .

In the following, we consider a concrete situation in which (a) the size of the spatial population is unknown to the sampler (there is no sampling frame) and (b) the inclusion probabilities are also unknown to the sampler (the sample is not the result of applying a probability sampling design). In the absence of auxiliary variables that would be (strongly) correlated with y and exhaustively known on  $\mathcal{D}$ , the only data available are the  $y_i$  values for  $i \in s$  and the sample size  $n_s$ . Thus, by default, the population mean can be estimated by the sample mean  $\overline{y}_s$ , and the p-variance of  $\overline{y}_s$  can be estimated using:

$$\widehat{\mathcal{V}}_{p}\left(\overline{y}_{s}\right) = \frac{s_{y}^{2}}{n_{s}} \tag{1}$$

#### <sup>150</sup> 3. Sampling process models

The fact that only a part (i.e., the sample) of the population of spatial sampling units is considered for data collection represents a source of uncertainty that may be modeled. This can be referred to as a *sampling model* (Hobbs and Hooten, 2015, Sec. 1.1.3) or an *inclusion model* (Gelman et al., 2014, Sec. 8.2), although both terms are polysemous which can lead to confusion.

As noted by Hájek (1964, p. 1492), any probability distribution p(s) can be used as a mathematical model for any sampling procedure, experiment, or method. Thus, in this article, we model a sampling process using an appropriate probability sampling design (Aubry, 2023, Sec. 1.2.4). Basic sampling processes (without replacement) can be modeled by one of four one-stage sampling designs, described below, depending on whether the propensities are equal or unequal and on whether the sample size is fixed or variable (Fig. 2).

![](_page_8_Figure_10.jpeg)

Figure 2: The four one-stage probability sampling designs used as without-replacement sampling process models. POISS-WOR: Poisson sampling. BERNWOR: Bernoulli sampling. CPS: Conditional Poisson sampling. SRSWOR: Simple random sampling. The sampling designs are derived from each other by asymmetric relations of equalization of inclusion probabilities  $\pi_i$  ( $i \in U$ ) or conditioning on a fixed sample size n.

#### <sup>161</sup> 3.1. Preferential sampling — variable size (POISSWOR)

When the sample size is variable and the propensities vary across sampling units, the sampling process can be modeled by Poisson sampling, which we denote as POISSWOR.

In Poisson sampling, (i) the inclusion of a unit  $i \in \mathcal{U}$  in sample *s* is governed by a Bernoulli trial with probability  $\pi_i$  and (ii) the resulting random sample size  $n_s$  follows a Poisson binomial distribution  $(n_s \sim \text{PoiBin}(N, \pi))$  (Hájek, 1981, Ch. 6; Särndal et al., 1992, Sec. 3.5; Tillé, 2006, Sec. 5.5; Tillé, 2020, Sec. 5.8). Let us note that the Poisson binomial distribution is here the generalization of the binomial distribution for Bernoulli trials with unequal probabilities (see, e.g., Wang, 1993) and should not be confused with the compound distribution described by Johnson et al. (2005, Sec. 9.5). A necessary condition for using POISSWOR as a model for variable-size sampling processes without replacement is that unit inclusions must be assumed to result from independent events. This assumption is broadly consistent with the situation in which the sample results from the actions of many people who do not coordinate with each other

<sup>173</sup> many people who do not coordinate with each other.

#### 174 3.2. Nonpreferential sampling — variable size (BERNWOR)

<sup>175</sup> When the sample size is variable and the propensities are sufficiently similar to each other such <sup>176</sup> that they can be considered equal to a constant probability  $\pi$ , the sampling process can be modeled <sup>177</sup> by Bernoulli sampling (also known as *binomial sampling*), which we denote as BERNWOR.

In Bernoulli sampling, (i) the inclusion of any unit in the sample is governed by a Bernoulli trial with probability  $\pi$  and (ii) the resulting variable sample size  $n_s$  follows a binomial distribution  $(n_s \sim \text{Bin}(N,\pi))$  (Särndal et al., 1992, Sec. 3.2; Tillé, 2006, Sec. 4.3; Tillé, 2020, Sec. 3.2).

BERNWOR is a special case of POISSWOR when the probabilities  $\pi_i$   $(i \in \mathcal{U})$  are equal (Fig. 2).

<sup>182</sup> When modeling a sampling process, BERNWOR has the same independence condition as POISSWOR.

#### <sup>183</sup> 3.3. Preferential sampling — fixed size (CPS)

Conditional on a fixed sample size  $(n_s = n \text{ for all } s)$ , POISSWOR becomes conditional Poisson sampling, abbreviated as CPS (Hájek, 1981, Ch. 14; Tillé, 2006, Sec. 5.6; Tillé, 2020, Sec. 5.9) (Fig. 2). This is a general model suitable for a fixed-size preferential sampling process.

### <sup>187</sup> 3.4. Nonpreferential sampling — fixed size (SRSWOR)

Conditional on a fixed sample size  $(n_s = n \text{ for all } s)$ , BERNWOR becomes simple random sampling without replacement, abbreviated as SRSWOR, for which  $\pi = n/N$  (Särndal et al., 1992, Sec. 3.3.1; Tillé, 2006, Sec. 4.4; Tillé, 2020, Sec. 3.3) (Fig. 2). SRSWOR is also a special case of CPS when the probabilities  $\pi_i$  ( $i \in U$ ) are equal (Fig. 2).

Let us recall that in this article, we assume that N is unknown to the sampler, so the propensity  $\pi$  is also unknown. SRSWOR is an appropriate model for a fixed-size sampling process with approximately constant propensities.

### 195 3.5. Estimators

<sup>196</sup> Unlike the concrete situation we considered in Section 2.4, in Monte Carlo simulations, the  $\pi$ -vector <sup>197</sup> is known, as is the population size N. In this case, the population mean can be estimated without <sup>198</sup> p-bias using the expansion estimator for the population total (Horvitz-Thompson estimator; see, e.g., <sup>199</sup> Hedayat and Sinha, 1991, Ch. 3) to form the following weighted estimator:

$$\overline{y}_{\pi} = \frac{1}{N} \sum_{i \in s} \frac{y_i}{\pi_i} \tag{2}$$

<sup>200</sup> For BERNWOR and SRSWOR, the estimator (2) simplifies to:

$$\overline{y}_{\pi} = \frac{1}{N} \frac{1}{\pi} \sum_{i \in s} y_i \tag{3}$$

which for SRSWOR gives the identity  $\overline{y}_{\pi} = \overline{y}_s$  since, in this case, we have  $\pi = n/N$ .

<sup>202</sup> For sampling processes modeled by BERNWOR or POISSWOR (variable sample sizes), the units are

<sup>203</sup> included in the sample independently. This gives a fairly simple expression for the *p*-unbiased estimator

of the sampling variance (*p*-variance) of the weighted estimator (2) (see Särndal et al., 1992, p. 289).

<sup>205</sup> The expression for POISSWOR is written as:

$$\widehat{\mathcal{V}}_{p}\left(\overline{y}_{\pi}\right) = \frac{1}{N^{2}} \left[ \sum_{i \in s} \frac{1 - \pi_{i}}{\pi_{i}^{2}} y_{i}^{2} \right]$$

$$\tag{4}$$

<sup>206</sup> which simplifies for BERNWOR to:

$$\widehat{\mathcal{V}}_{p}\left(\overline{y}_{\pi}\right) = \frac{1}{N^{2}} \left[ \frac{1-\pi}{\pi^{2}} \sum_{i \in s} y_{i}^{2} \right]$$
(5)

<sup>207</sup> In the case of fixed-size designs, in general, to estimate without bias the *p*-variance, it is necessary to

know the joint probabilities  $\pi_{ij} > 0$  that two units  $i \neq j \in \mathcal{U}$  are simultaneously included in the sample.

The general expression for the *p*-unbiased estimator of the *p*-variance of the weighted estimator (2)

for an unequal-probability, fixed-size, without-replacement sampling design such as the CPS is (see Sen-Yates-Grundy estimator, Hedayat and Sinha, 1991, p. 52, Eq. 3.16):

$$\widehat{\mathcal{V}}_{p}\left(\overline{y}_{\pi}\right) = \frac{1}{N^{2}} \left[ \sum_{(i < j) \in s} \frac{\pi_{i}\pi_{j} - \pi_{ij}}{\pi_{ij}} \left(\frac{y_{i}}{\pi_{i}} - \frac{y_{j}}{\pi_{j}}\right)^{2} \right]$$
(6)

which simplifies for SRSWOR since then we have  $\pi_i = n/N$  for all  $i \in \mathcal{U}$  and  $\pi_{ij} = n(n-1)/[N(N-1)]$ for all  $i \neq j \in \mathcal{U}$ . After simplification, we obtain (e.g., Tillé, 2020, p. 29, Result 3.1):

$$\widehat{\mathcal{V}}_{p}\left(\overline{y}_{\pi}\right) = \left(1 - \pi\right) \frac{s_{y}^{2}}{n} \tag{7}$$

In the class of fixed-size without-replacement sampling designs with inclusion probabilities  $\pi$ , the CPS is the sampling design of maximum entropy (Hájek, 1981; Tillé, 2006; Tillé, 2020, Sec. 5.9). It follows that an approximation to the *p*-variance estimator (6) may be obtained without involving the joint inclusion probabilities  $\pi_{ij}$  ( $i \neq j \in \mathcal{U}$ ). We refer the reader to Tillé (2020, Sec. 5.14) and to the references cited by Aubry (2023, Sec. 1.3.2).

#### 219 4. Status assessment

From an operational perspective, for a given spatial domain  $\mathcal{D}$ , status at a given point in time 220 corresponds to either (i) the state of a variable of interest or (ii) the value of an indicator summarizing 221 the state of the variable of interest. In practice, the point in time is actually a time interval — ideally, 222 the shortest possible — which we have denoted  $\mathcal{T}$ . In the case of abundance, for example, at a certain 223 infra-annual time interval  $\mathcal{T}$  over which a biological population may be assumed to be (approximately) 224 closed both geographically and demographically, status in sense (i) may correspond to the spatial 225 distribution of abundance, while in sense (ii), it may refer to average abundance. In the following, we 226 refer to status in sense (ii) (indicator of interest). 227

#### 228 4.1. Status example

As an example of a status to be estimated over a  $\mathcal{D} \times \mathcal{T}$  spatiotemporal domain, we consider here the average species richness for a group of L species and a finite statistical population  $\mathcal{U}$  of N equal-size spatial sampling units, discretizing all or a part of the spatial domain  $\mathcal{D}$ . The variable of interest y is therefore species richness, and the indicator is the finite population mean  $\overline{y}_{\mathcal{U}}$  (the time interval  $\mathcal{T}$  is implicit and omitted from the notation).

In this example, spatial units are considered habitat patches that are disjunct within the spatial domain of interest (Wiens, 1976; Hall et al., 1997; Girvetz and Greco, 2007). A habitat patch may be included in the sample from a previously established list of all patches (sampling frame) that has been established by habitat mapping, in which case N is known. In the absence of a sampling frame, habitat patches are included in the sample based on the sampler's prior knowledge or on what the sampler encounters in the field, in which case N is generally unknown. This second case corresponds to the concrete situation of preferential sampling considered in this article (Section 2.4).

In the case of (inherently) nonpreferential sampling, units are selected with the same propensity  $\pi$ . With preferential sampling, units are selected based on a subjective assessment of their expected species richness, with propensities more or less positively correlated with y-values.

#### 244 4.2. Population model example

As an illustrative example of a variable of interest y, we consider species richness with values 245 ranging from 1 to L = 25, distributed according to a zero-truncated beta-binomial distribution 246 BetaBinZT(L,  $\alpha$ ,  $\beta$ ) with shape parameters  $\alpha = 1$  and  $\beta = 5$ . These values are realistic since they 247 are of the same order of magnitude as those obtained for the species richness of a set of 25 waterbird 248 species (ducks, geese, swans, coots, waders and grebes) breeding in European France (see the LIMAT 249 scheme described in Aubry et al., 2023, Fig. 1). According to the typology proposed by Aubry and 250 Francesiaz (2022, Section 2.4, Table 1), this statistical distribution is a type IV superpopulation model; 251 that is, it does not include information on spatial structure or covariates (auxiliary variables). This 252 model specifies an infinite set of values, but we are interested here in sampling a finite population 253 containing  $N = 5\,000$  spatial sampling units. 254

Among the infinite set of populations containing  $N = 5\,000 \text{ y-values following a BetaBinZT}(25, 1, 5)$ distribution, we choose one such that  $|\hat{\alpha} - \alpha| < 10^{-3}$  and  $|\hat{\beta} - \beta| < 10^{-3}$ . The estimator used for the shape parameters  $\alpha$  and  $\beta$  of the zero-truncated beta-binomial distribution is the second estimator proposed by Tripathi et al. (1994, Sec. 3.1). We can visually check that the histogram of the variable

 $_{259}$  y associated with the finite population  $\mathcal{U}$  we used is very close to that of the superpopulation model

(Fig. 3). Thus, we have  $\overline{y}_{\mathcal{U}} = 5.0062$  for comparison with its expected value in the model  $E_{\xi}(\overline{y}_{\mathcal{U}}) = 5.0062$ 

![](_page_11_Figure_3.jpeg)

Figure 3: Distribution of the superpopulation model (in dark green) and histogram of the simulated population for  $N = 5\,000$  (in light green).

#### 261 4.3. Monte Carlo study

In this section, we examine the estimation of  $\overline{y}_{\mathcal{U}}$  by  $\overline{y}_s$  in four situations, combining the preferential vs. nonpreferential nature of the sampling process and a variable vs. fixed sample size.

In the case of preferential sampling, the propensities are defined as being proportional to a size variable z ( $z_i > 0$  for all  $i \in \mathcal{U}$ ):

$$\pi_i = P \frac{z_i}{Z} \quad \text{with} \quad P = \sum_{i \in \mathcal{U}} \pi_i \quad \text{and} \quad Z = \sum_{i \in \mathcal{U}} z_i$$
(8)

In the case of POISSWOR or BERNWOR, we have  $P = E_p(n_s)$  (variable sample size); in the case of CPS or SRSWOR, we have P = n (fixed sample size).

Since we must respect the constraint  $\pi_i \leq 1$ , it follows that if we have  $z_i > Z/P$  for at least one unit, we must perform the preprocessing described, for example, by Aubry (2023, Sec. 3.1) (see also Tillé, 2020, Sec. 5.2). From the above, we deduce that  $R_{\pi y} = R_{zy}$  if and only if  $z_i \leq Z/P$  for all  $i \in \mathcal{U}$ . In the following, we consider cases where this condition holds (for a counterexample, see Aubry et al., 2020, Sec. 4.4).

We define a model for the size variable z in which the correlation with y is  $\rho_{zy}$ . We vary  $\rho_{zy}$ between 0 and 1 by attenuating perfect positive correlation — for correlation attenuation; see, e.g., Charles (2005) — by adding a Gaussian error term  $\epsilon \sim N(0, \sigma_{\epsilon}^2)$  to y. Thus, for each  $i \in \mathcal{U}$ :

$$z_{i} = \epsilon_{i} \qquad \text{with } \sigma_{\epsilon}^{2} > 0 \qquad \text{for } \rho_{zy} = 0$$
  

$$z_{i} = y_{i} + \epsilon_{i} \qquad \text{with } \sigma_{\epsilon}^{2} = S_{y}^{2}(1 - \rho_{zy}^{2})/\rho_{zy}^{2} \qquad \text{for } 0 < \rho_{zy} < 1$$
  

$$z_{i} = y_{i} \qquad \text{with } \sigma_{\epsilon}^{2} = 0 \qquad \text{for } \rho_{zy} = 1$$
(9)

Ultimately, to guarantee  $z_i > 0$  (and hence,  $\pi_i > 0$ ) for all  $i \in \mathcal{U}$ , we apply the linear transformation  $z_{ii} - \min(\mathbf{z}) + 0.1$ .

In the same spirit as obtaining  $\mathbf{y}$ , we want the value of the finite population correlation  $R_{zy}$ to approximately match the superpopulation correlation  $\rho_{zy}$ . To achieve this, we generate random outcomes of the  $\mathbf{z}$ -vector until we obtain  $|R_{zy} - \rho_{zy}| < 10^{-6}$ . Having obtained the  $\mathbf{z}$ -vector, we then set a value for P and compute the  $\boldsymbol{\pi}$ -vector according to expression (8).

In the following, we replicate a given sampling process model with P = 500, i.e., with 10% as the sampling fraction (fixed sample size) or expectation of the sampling fraction (variable sample size). For each replication, we compute  $\overline{y}_s$  and  $\widehat{V}_p(\overline{y}_s)$ . We use  $10^6$  replications to accurately approximate the *p*-distributions of these two estimators. Therefore, we can assess the possible estimation bias of the population mean  $\overline{y}_{\mathcal{U}}$  by the sample mean  $\overline{y}_s$  and also that of the *p*-variance  $V_p(\overline{y}_s)$  by the default estimator  $\widehat{V}_p(\overline{y}_s)$  (Eq. 1).

To evaluate the estimation bias of a parameter  $\omega$  by using an estimator  $\hat{\omega}$ , we define a *bias index* BI = E( $\hat{\omega}$ )/ $\omega$ . The bias E( $\hat{\omega}$ ) –  $\omega$  is positive if BI > 1, zero for BI = 1, and negative if BI < 1. Here, to assess the estimation bias for the *p*-variance of the sample mean, we form:

$$BI = \frac{E_{MC} \left( \widehat{V}_{p} \left( \overline{y}_{s} \right) \right)}{V_{MC} \left( \overline{y}_{s} \right)}$$
(10)

where  $E_{MC}(\cdot)$  and  $V_{MC}(\cdot)$  are the mean and variance calculated over the 10<sup>6</sup> replications of the Monte Carlo study.

- <sup>293</sup> 4.3.1. Nonpreferential sampling variable size (BERNWOR)
- The sampling process model corresponding to (inherently) nonpreferential sampling of variable size is BERNWOR. The *p*-distribution of  $\overline{y}_s$  for  $\pi = 0.1$  (E<sub>p</sub> ( $n_s$ ) = 500) is shown in Fig. 4.a.

In the BERNWOR case, the *p*-unbiased estimator of the population mean is the weighted estimator  $\bar{y}_{\pi}$ , which differs from the sample mean  $\bar{y}_s$ . This difference implies that the estimation of the population mean using the sample mean is *p*-biased. However, the bias is generally negligible and of no consequence, as shown in our example (Fig. 4.a). Here, although mathematically biased, the sample mean is actually a preferable estimator to the weighted estimator (see Särndal et al., 1992, p. 64, Eq. 3.2.6 and Tillé, 2020, p. 35).

The default estimator for the *p*-variance of  $\overline{y}_s$  (Eq. 1) is slightly positively biased since we obtain BI  $\approx$  1.1136. The *p*-variance of  $\overline{y}_s$  under the BERNWOR model is approximately the same as that under the SRSWOR model (Särndal et al., 1992, p. 65) (see Section 4.3.2 and Fig. 4).

305 4.3.2. Nonpreferential sampling — fixed size (SRSWOR)

The sampling process model corresponding to the (inherently) nonpreferential fixed-size sampling is SRSWOR. The *p*-distribution of  $\overline{y}_s$  for  $\pi = 0.1$  (n = 500) is shown in Fig. 4.b.

As recalled in Section 3, in the SRSWOR case, we have the identity  $\overline{y}_{\pi} = \overline{y}_s$ . This identity implies that the estimation of the population mean using the sample mean is *p*-unbiased (Fig. 4.b).

Conversely, the default estimator of the p-variance (Eq. 1) is positively biased since we obtain 310  $BI \approx 1.1111$ . This is because the unbiased estimator in the SRSWOR model (Eq. 7) introduces a 311 finite population correction (see Cochran, 1977, Sec. 2. 6) fpc =  $1 - \pi$ , which is closer to 0 the closer 312  $\pi$  is to 1. At the extreme, the p-variance becomes zero when n = N (finite population consistency; 313 see Cochran, 1977, Sec. 2.4, p. 21 or Hankin et al., 2019, p. 325). The magnitude of the p-variance 314 overestimation depends on the value of the sampling fraction  $\pi = n/N$ , which remains unknown when 315 N is not known. Let us note that the value obtained by Monte Carlo simulation for BI agrees with 316 the theoretical value  $(1 - \pi)^{-1}$  for  $\pi = 0.1$ . 317

![](_page_12_Figure_12.jpeg)

Figure 4: Approximate sampling distributions using  $10^6$  replications of each nonpreferential sampling model for  $\pi = 0.1$ . (a) BERNWOR. (b) SRSWOR. The population mean is shown by the red dashed line. The average of the values taken by  $\overline{y}_s$  is shown by the blue line.

#### 318 4.3.3. Preferential sampling — variable size (POISSWOR)

The sampling process model corresponding to variable-size preferential sampling is POISSWOR. The vector of probabilities  $\boldsymbol{\pi}$  is computed as explained in Section 4.3 for  $E_p(n_s) = 500$  (variable sample size). The *p*-distributions of  $\overline{y}_s$  for  $R_{\pi y} = 0.5$  and  $R_{\pi y} = 0.9$  are shown in Fig. 5.

![](_page_13_Figure_0.jpeg)

Figure 5: Approximate sampling distributions using  $10^6$  replications of the POISSWOR model with P = 500. (a)  $R_{\pi y} = 0.5$ . (b)  $R_{\pi y} = 0.9$ . The population mean is shown by the red dashed line. The average of the values taken by  $\overline{y}_s$  is shown by the blue line.

In the case of POISSWOR, the *p*-unbiased estimator of the population mean is the weighted estimator  $\bar{y}_{\pi}$ , which is different from the sample mean  $\bar{y}_s$ . Therefore, like with BERNWOR, the estimation of the population mean using the sample mean is *p*-biased. However, unlike the BERNWOR case, this should not be neglected since the relative *p*-bias is approximately 10.2% for  $R_{\pi y} = 0.5$  and 27.6% for  $R_{\pi y} = 0.9$  (Fig. 5). The default estimator for the *p*-variance of  $\bar{y}_s$  (Eq. 1) is positively biased since we obtain BI  $\approx 1.1411$  for  $R_{\pi y} = 0.5$  and BI  $\approx 1.1788$  for  $R_{\pi y} = 0.9$ .

328 4.3.4. Preferential sampling — fixed size (CPS)

The sampling process model corresponding to fixed-size preferential sampling is CPS. The  $\pi$ -vector is the same as that for POISSWOR. The *p*-distributions of  $\overline{y}_s$  for  $R_{\pi y} = 0.5$  and  $R_{\pi y} = 0.9$  are shown in Fig. 6.

![](_page_13_Figure_5.jpeg)

Figure 6: Approximate sampling distributions using  $10^6$  replications of the CPS model for P = 500. (a)  $R_{\pi y} = 0.5$ . (b)  $R_{\pi y} = 0.9$ . The population mean is shown by the red dashed line. The average of the values taken by  $\overline{y}_s$  is shown by the blue line.

In our example, the case of CPS is very similar to that of POISSWOR since the relative *p*-bias of the sample mean is approximately 10.2% for  $R_{\pi y} = 0.5$  and 27.6% for  $R_{\pi y} = 0.9$  (Fig. 6). The default estimator for the *p*-variance of  $\overline{y}_s$  (Eq. 1) is also positively biased since we obtain BI  $\approx 1.1387$  for  $R_{\pi y} = 0.5$  and BI  $\approx 1.1793$  for  $R_{\pi y} = 0.9$ , values that are close to those obtained in the POISSWOR case.

#### <sup>337</sup> 4.3.5. Zero-correlation between the propensity and the variable of interest

Thus far, we have simulated preferential sampling for  $R_{\pi y} = 0.5$  and  $R_{\pi y} = 0.9$ . In this section, we examine the degenerate situation where  $R_{\pi y} = 0$ .

When the correlation between the propensities and y-values is zero, the relative p-bias of the sample mean is essentially zero for both POISSWOR and CPS (Fig. 7). The default estimator for the p-variance of  $\overline{y}_s$  (Eq. 1) is positively biased, with BI  $\approx 1.1223$  for the POISSWOR and BI  $\approx 1.1201$  for the CPS. Although the propensities are variable, sampling is nonpreferential (Fig. 7). The situation is broadly equivalent to that encountered with the BERNWOR and SRSWOR models (Fig. 4).

![](_page_14_Figure_1.jpeg)

Figure 7: Approximate sampling distributions using  $10^6$  replications of each unequal propensity sampling model for P = 500 and  $R_{\pi y} = 0$ . (a) POISWOR. (b) CPS. The population mean is shown by the red dashed line. The average of the values taken by  $\overline{y}_s$  is shown by the blue line.

### 345 4.4. An analytical formula for the bias of the sample mean

In this section, we focus on the formal expression of the *p*-bias of the sample mean  $\overline{y}_s$  for estimating the population mean  $\overline{y}_{\mathcal{U}}$ . For algebraic details, the reader is referred to Appendix B.

For a sampling design of variable size, the p-bias of the sample mean can be written as (Appendix B.1):

$$B_{p}(\overline{y}_{s}) = E_{p}(\overline{y}_{s}) - \overline{y}_{\mathcal{U}} \approx \frac{R_{\pi y} S_{\pi} S_{y}}{\overline{\pi}_{\mathcal{U}}}$$
(11)

where  $R_{\pi y}S_{\pi}S_y = S_{\pi y}$  is the (adjusted) covariance between  $\pi$  and  $\mathbf{y}$  (see Appendix A).

In our example, we have  $\overline{y}_{\mathcal{U}} \approx 5$ ,  $S_y \approx 3.7787$ ,  $N = 5\,000$  and  $\overline{\pi}_{\mathcal{U}} = 0.1$ . For  $R_{\pi y} = 0.5$ , we have  $S_{\pi} \approx 0.0272$ , from which we obtain  $B_p(\overline{y}_s) \approx 0.51$  and a relative *p*-bias of approximately 10.2%, in agreement with what we obtained via Monte Carlo simulation in the POISSWOR case (Section 4.3.3). For  $R_{\pi y} = 0.9$ , we have  $S_{\pi} \approx 0.0407$ , from which we obtain  $B_p(\overline{y}_s) \approx 1.38$  and a relative *p*-bias of approximately 27.6%, which again matches the value found earlier by Monte Carlo simulation.

In the case of a fixed-size sampling design, the *p*-bias of the sample mean admits an exact expression (Eq. B.11, Appendix B.2), but we can also use expression (11) as a close approximation when N is not too small. Consequently, we obtain the same results for CPS as for POISSWOR.

As expected, the observations in the cases studied by Monte Carlo simulation match the results computed by using formula (11). In the context of Monte Carlo simulation, when we are interested in the *p*-bias of the sample mean, this formula allows us to avoid replicating a sampling process model. We use this computational shortcut when appropriate throughout the remainder of this article.

Examination of formula (11) shows that the sample mean is a *p*-unbiased estimator of the population 363 mean in the following two situations: (i) when the propensities are variable  $(S_{\pi} > 0)$ , if there is no 364 correlation between the propensities and the y-values  $(R_{\pi y} = 0)$  or (ii) when propensities do not vary 365  $(S_{\pi}=0)$ , since then the covariance and correlation are zero  $(S_{\pi}=0 \Rightarrow S_{\pi y}=0 \Rightarrow R_{\pi y}=0)$ . If the 366 covariance between the propensities and the y-values remains unchanged (fixed  $S_{\pi y}$ ), the lower the 367 sampling fraction (fixed sample size) or its expectation (variable sample size)  $\overline{\pi}_{\mathcal{U}}$  is, the greater the 368 bias. As expected, the p-bias is zero in the case of a census since the propensities do not vary (and are 369 all equal to 1), which is the same as in case (ii) above. 370

#### **5.** Trend assessment

For operational purposes, a trend can be defined as a pattern of change in the indicator of interest over time. In the following, we consider a monitoring program designed to study the trend of a population parameter over a period  $\Delta$ . The monitoring program starts at time  $t_0$  and ends at time  $t_0 + \Delta$ .

#### 376 5.1. Trend example

As an example of a trend to be estimated, we consider here the trend in the abundance of a species over a bounded domain  $\mathcal{D}$  of the Euclidean plane, partitioned by a population  $\mathcal{U}$  of subdomains  $d_i$ (i = 1, 2, ..., N), formally:

$$\mathcal{D} = \bigcup_{i \in \mathcal{U}} d_i \quad \text{with } d_i \cap d_j = \emptyset \quad \text{for } i \neq j \in \mathcal{U} \quad \text{and} \quad |\mathcal{D}| = \sum_{i \in \mathcal{U}} |d_i|$$
(12)

In a general sense, subdomains can have different shapes and sizes. To simplify the presentation, in what follows, we consider the case where  $\mathcal{D}$  is an iso-oriented rectangular domain (i.e., its sides are parallel to the abscissa and ordinate axes), lying between  $x_{\min}$  and  $x_{\max}$  on the abscissa and between  $y_{\min}$  and  $y_{\max}$  on the ordinate. With such a domain, subdomains can simply be square cells of equal size. Thus,  $\mathcal{D}$  is discretized into a spatial population of  $NX \times NY = N$  grid cells of equal size, where NX and NY are the numbers of columns and rows in the grid, respectively.

For the species of interest, at time t (omitted from the notation for simplicity), an individual has a certain probability density f(u) > 0 of being present at a point  $u \in \mathcal{D}$  with Cartesian coordinates  $(x_u, y_u)$ . The value of the variable of interest  $y_i$  is the number of individuals present in grid cell  $d_i$  $(i \in \mathcal{U})$ , i.e., the number of points u where the species is present such that  $u \in d_i$  (here, a point represents only one individual).

When the size N of the sampled spatial population  $\mathcal{U}$  is known, it is the same, to within a factor i.e., N for the point estimator and  $N^2$  for its p-variance — to estimate the mean or the total abundance. For the concrete situation of preferential sampling that we address in this article (Section 2.4), we consider that N is unknown to the sampler. Thus, in what follows, the parameter of interest is the population mean  $\overline{y}_{\mathcal{U}}$ , estimated by the sample mean  $\overline{y}_s$ , as in Section 4. We assume that the spatial variability of the variable of interest is greater on the abscissa than on the ordinate and that at time  $t_0$ , there is a spatial gradient in abundance from  $x_{\min}$  to  $x_{\max}$ .

#### <sup>398</sup> 5.2. Population model example

<sup>399</sup> 5.2.1. Simulating an inhomogeneous binomial point process

The spatial distribution of a fixed number of individuals M over  $\mathcal{D}$  can be modeled by an inhomogeneous binomial point process (abbreviated as IBPP) with a spatial intensity function  $\lambda(u)$ . Since a binomial point process is a Poisson point process conditional on M, simulation of an IBPP can be performed as for an inhomogeneous Poisson point process using the Lewis-Shedler method (see, e.g., Illian et al., 2008, p. 119; Baddeley et al., 2016, Sec. 5.4.2). An IBPP is simulated as follows:

- 405 1. Generate a point  $u \in \mathcal{D}$  following a Bernoulli point process.
  - 2. The probability of keeping this point is computed as  $p(u) = \lambda(u)/\lambda^*$ , where  $\lambda^*$  is the maximum value of the intensity on  $\mathcal{D}$ :

$$\lambda^* = \max_{u \in \mathcal{D}} \lambda(u)$$

- 3. A Bernoulli trial is performed with probability p(u). If the trial is successful, then the point u is kept.
- 408 4. Steps 1 to 3 are repeated until M points are drawn.

From a realization of the IBPP, the **y**-vector is obtained by counting the number of points within each grid cell  $d_i$  ( $i \in \mathcal{U}$ ). Simulating the realizations of the variable of interest directly by allocating the M individuals among the N grid cells is equivalent to doing so. First, for i = 1, 2, ..., N, the probability  $p_i$  of drawing grid cell  $d_i$  is computed as follows (we recall that the grid cells form a partition of  $\mathcal{D}$ ):

$$p_{i} = \frac{\int_{d_{i}} \lambda(u) du}{\sum_{i \in \mathcal{U}} \int_{d_{i}} \lambda(u) du} = \frac{\int_{d_{i}} \lambda(u) du}{\int_{\mathcal{D}} \lambda(u) du}$$
(13)

In our case, for all  $i \in \mathcal{U}$ , we have an area  $|d_i|$  (i) that is very small compared to area  $|\mathcal{D}|$  and (ii) that is a constant (the square grid cells have the same area). Therefore, in two steps, we obtain the following approximation:

$$p_i \approx \frac{\lambda(u_i) |d_i|}{\sum_{i \in \mathcal{U}} \lambda(u_i) |d_i|} = \frac{\lambda(u_i)}{\sum_{i \in \mathcal{U}} \lambda(u_i)}$$
(14)

- 417 where  $u_i$  is the barycenter of grid cell  $d_i$ .
- $_{418}$  The *M* individuals are allocated as follows:
- 419 1. Initialize  $y_i \leftarrow 0$  for i = 1, 2, ..., N.
- 2. Select a grid cell of index j by unequal probability sampling with replacement (also known as multinomial sampling) (e.g., Johnson et al., 1997, Sec. 8; Aubry, 2023, Remark 6) with drawing probabilities  $p_i$  (i = 1, 2, ..., N).
- 423 3. Increment  $y_j \leftarrow y_j + 1$ .
- 4. Steps 2 and 3 are repeated until M grid cells are drawn.
- <sup>425</sup> 5.2.2. Spatial intensity function

As mentioned above, we assume that the spatial distribution of individuals over  $\mathcal{D}$  is governed by a spatial intensity function  $\lambda(x)$  describing a one-dimensional gradient along the x-axis (we now use x instead of  $x_u$  to simplify the notation). The shape of this spatial gradient is determined by the shape parameter  $\eta$  (Fig. 8) according to the expression:

$$\lambda(x) = \frac{\lambda^* \exp(\eta \, x')}{\delta} \qquad \text{with} \quad x \in [x_{\min}, x_{\max}] \tag{15}$$

430 where  $x' \in [0, 1]$  is defined as follows:

$$x' = \frac{x - x_{\min}}{x_{\max} - x_{\min}} \tag{16}$$

431 and

$$\delta = \begin{cases} 1 & \text{if } \eta < 0\\ \exp(\eta) & \text{otherwise} \end{cases}$$
(17)

The degenerate case  $\lambda(x) = \lambda^*$  is obtained for  $\eta = 0$  (Fig. 8) and corresponds to a homogeneous binomial point process and equiprobable allocation.

![](_page_16_Figure_16.jpeg)

Figure 8: Plot of the intensity function  $\lambda(x)$  on the interval  $x \in [0, 1]$  with  $\lambda^* = 10$ , for  $\eta = 0, 1, 2, 3, 4$ .

We assume that the shape of the spatial gradient changes over time in such a way that its direction is completely reversed at time  $t_0 + \Delta$ , i.e., it is ultimately from  $x_{\text{max}}$  to  $x_{\text{min}}$ . This can be achieved with the shape parameter  $\eta$  (Eq. 15) by gradually decreasing its initial value  $\eta_0$  at time  $t_0$  until  $\eta_{\Delta} = -\eta_0$ 

437 at time  $t_0 + \Delta$  (example in Fig. 9).

![](_page_17_Figure_2.jpeg)

Figure 9: Plot of the intensity function  $\lambda(x)$  on the interval  $x \in [0, 1]$  with  $\lambda^* = 10$ , at time  $t_0$  ( $\eta_0 = 4$ , solid line) and at time  $t_0 + \Delta$  ( $\eta_{\Delta} = -4$ , dashed line).

#### 438 5.3. Monte Carlo study

We assume that the number of individuals M in  $\mathcal{D}$  remains constant over the monitoring period (no demographic change).

Preferential sampling of fixed size n is performed according to the sampler's view of the variation 441 in abundance over  $\mathcal{D}$  at time  $t_0$ , favoring the sampling units expected to be the richest in individuals, 442 i.e., those with the highest y-values. As mentioned in Section 3, this type of preferential sampling 443 process can be modeled by conditional Poisson sampling (CPS). Under preferential sampling, we have 444 shown that using the sample mean biases the mean abundance estimation over  $\mathcal{D}$  (Section 4). This is 445 not a problem for assessing trends or changes between two points in time, as long as the bias remains 446 (approximately) constant over the monitoring period, which here implies that the units that were 447 expected to be the richest in individuals at time  $t_0$  remain so between  $t_0$  and  $t_0 + \Delta$ . Conversely, 448 let us consider a directional change in the spatial distribution of individuals in  $\mathcal{D}$  over the monitoring 449 period. In this situation, using the sample mean to estimate the population mean results in estimating 450 a spurious trend. 451

In our example, at the start of the monitoring program, the sampler knows that there is a spatial gradient in abundance from  $x_{\min}$  to  $x_{\max}$  but does not know its exact shape. By default, we simulate this situation by using a size variable expressed as  $z_i = a \times x_i + b$  with (a, b) such that  $z_i > 0$  for all  $i \in \mathcal{U}$  to guarantee a probability  $\pi_i > 0$  of being part of the sample.

We thus have (i) a fixed vector of propensities ( $\boldsymbol{\pi}$ ) computed from the size variable ( $\mathbf{z}$ ) (Eq. 8, Fig. 456 10, top panel) and (ii) a random vector of abundances  $(\mathbf{y})$  obtained according to the intensity function 457  $\lambda(x)$  (Eq. 15, Fig. 10, bottom panel). For a given realization of the spatial variation model defined by 458  $\lambda(x)$ , the correlation between the propensities and the abundances takes a certain fixed value  $R_{\pi y}$ . The 459 model can generate an infinite number of y-vectors, hence an infinite number of  $R_{\pi y}$ -values (within the 460 range of variation of  $R_{\pi y}$  for this model). If the population size is sufficiently large, for convenience, 461 the correlation between the propensities and the abundances in the model  $(\rho_{\pi y})$  can be assimilated to 462 the  $\xi$ -expectation of the population correlation  $R_{\pi y}$  (Appendix A, Fig. 10): 463

![](_page_18_Figure_0.jpeg)

Figure 10: Schematic representation of the relationships between the simulation components. Top panel: The size variable (**z**) determines the propensities ( $\pi$ ), which remain fixed. Bottom panel: The intensity function of the inhomogeneous binomial point process ( $\lambda(x)$ ) determines the vector of abundances (**y**), which is random. If the population size is sufficiently large, for convenience, the correlation in the model ( $\rho_{\pi y}$ ) can be assimilated to the  $\xi$ -expectation of the finite population correlation  $R_{\pi y}$ .

In the following, we first consider the case of a fixed population (i.e., a given realization  $\mathbf{y}$ ) and then the case of the superpopulation model itself. We decrease the shape parameter  $\eta$  to simulate the reversal

of the spatial gradient in abundance over  $\mathcal{D}$  between  $t_0$  and  $t_0 + \Delta$ , where we assume that  $\eta$  decreases

linearly with time. For simplicity, we decrease  $\eta$  by one unit per time unit (e.g., per year or decade).

<sup>468</sup> By varying the shape parameter from  $\eta_0 = 4$  at time  $t_0$  to  $\eta_{\Delta} = -4$  at time  $t_0 + \Delta$  (Fig. 9), we cover a

<sup>469</sup> monitoring period of  $\Delta = 8$  time units and have a (short) time series of 9 mean abundance estimates <sup>470</sup> to fit a (parametric) temporal trend model.

We simulate a simple situation where an iso-oriented rectangular domain  $\mathcal{D}$  is discretized by N = 2500 square cells organized according to a grid of NX = 100 columns and NY = 25 rows. For simplicity, we set  $x_{\min} = y_{\min} = 0$ ,  $x_{\max} = 100$  and  $y_{\max} = 25$  so that each grid cell has a unit area. The maximum intensity is  $\lambda^* = 10$  (the number of points per unit area, i.e., also per grid cell).

Regardless of the value of  $\eta$ ,  $M = 10\,000$  individuals are randomly allocated among the  $N = 2\,500$ grid cells, as explained in Section 5.2. Therefore, we have  $\overline{y}_{\mathcal{U}} = M/N = 4$ . To define the size variable  $z_i$ , we choose a = b = 0.1, that is,  $z_i = 0.1 \times x_i + 0.1$  ( $i \in \mathcal{U}$ ). We again use a sampling fraction of 10%, i.e., n = 250. The variable propensities that form the  $\pi$ -vector result from the previous choices regarding z and n; they remain constant throughout the simulations (Fig. 10, top panel).

#### 480 5.3.1. Fixed population

Like in Section 4, for each value of  $\eta$ , we keep a **y**-vector of abundances for which the correlation  $R_{\pi y}$  is approximately equal to a value of  $\rho_{\pi y}$  compatible with the value of  $\eta$ . For each case, we generate realizations of the **y**-vector until we obtain one for which we have  $|R_{\pi y} - \rho_{\pi y}| < 10^{-4}$ . We take  $\eta = 4, 1, 0, -1, -4$ . Compatible correlation values are  $\rho_{\pi y} = 0.8, 0.5, 0.0, -0.5, -0.8$ .

In the case of fixed populations, we are interested in the *p*-distribution of the sample mean  $(\overline{y}_s)$ and in the *p*-bias in particular. We replicate 10<sup>6</sup> times the conditional Poisson sampling (CPS) based on the  $\pi$ -vector to accurately approximate the *p*-distribution of  $\overline{y}_s$ . Several samples obtained by CPS are shown in Fig. 11. As expected for a (putative) linear gradient in abundance from  $x_{\min}$  to  $x_{\max}$ and preferential sampling, we see a greater spatial concentration of selected units toward the higher abscissas and a unit deficit toward the lower abscissas, with a continuum of unit densities in between.

![](_page_19_Figure_0.jpeg)

Figure 11: Several samples selected by CPS with inclusion probabilities proportional to a size variable z that varies linearly according to the equation  $z_i = 0.1 \times x_i + 0.1$   $(i \in \mathcal{U})$  and for n = 250. See the text for details.

In agreement with the analytical formula for the bias of the sample mean (Eq. 11), it appears that the bias is initially positive for  $R_{\pi y} > 0$ , vanishes for  $R_{\pi y} = 0$  and then becomes negative for  $R_{\pi y} < 0$  (Fig. 12). Sampling that planned to be preferential at the beginning of the monitoring program becomes nonpreferential and eventually antipreferential due to spatial gradient reversal.

![](_page_20_Figure_0.jpeg)

Figure 12: Results of the Monte Carlo study for the five fixed populations. (1)  $\eta = 4$ ,  $R_{\pi y} = 0.8$ . (2)  $\eta = 1$ ,  $R_{\pi y} = 0.5$ . (3)  $\eta = 0$ ,  $R_{\pi y} = 0$ . (4)  $\eta = -1$ ,  $R_{\pi y} = -0.5$ . (5)  $\eta = -4$ ,  $R_{\pi y} = -0.8$ . (a) Finite populations simulated on a  $100 \times 25$  grid as a function of the spatial variation model for the set of  $\eta$ -values. The marginal graph (in gray) represents the mean abundance in the grid cells along the x-axis. This reflects the realization of the spatial variation model. (b) Approximations of the *p*-distributions of the sample mean  $(\bar{y}_s)$ ;  $10^6$  samples are generated by conditional Poisson sampling. The finite population mean is represented by the red dashed line. The average of the values taken by  $\bar{y}_s$  is shown by the blue line. The figures are adjusted to maximize visibility, so the scales may differ from one figure to another. See the text for details.

In practice, a sample of spatial units s obtained at time  $t_0$  may remain unchanged over the monitoring period (permanent plots, fixed monitoring network). In the following, we assume that this is indeed the case. In our example of a reversal spatial gradient, with simulated preferential sampling and by using the sample mean  $\overline{y}_s$  as an estimator of the population mean  $\overline{y}_{\mathcal{U}}$ , there is a serious risk of systematically highlighting a downward trend, while in fact, the abundance in  $\mathcal{D}$  remains constant over the monitoring period (no demographic change).

To investigate this hypothesis, at time  $t_0$ , we simulate 10<sup>4</sup> samples using CPS. We compute  $\overline{y}_s$  for 501 each sample and for each of the abundance vectors at the times corresponding to the different values of 502  $\eta$ . This gives  $10^4$  time series of mean abundance estimates. We add four more  $\eta$  values than those used 503 above, with  $\eta = 3, 2, -2, -3$  for which compatible correlation values are  $\rho_{\pi y} = 0.79, 0.72, -0.72, -0.79$ . 504 With a time series of 9 mean abundance estimates at hand (for  $\eta = 4, 3, 2, 1, 0, -1, -2, -3, -4$ ), we fit 505 a linear regression model as a function of time. We use weighted least squares regression (see, e.g., 506 Press et al., 1989, Sec. 14.2) to account for the *p*-variance estimates. For each sample, we thus obtain 507 a straight line with slope m, and for all the  $10^4$  samples generated by CPS, we obtain an envelope of 508 straight lines reflecting the sampling variability of the trend highlighted by the monitoring program. 509 As expected, in our example, a downward trend is obtained in all the cases, with an average slope of 510

<sup>511</sup>  $E_p(m) \approx -0.56$  (Fig. 13).

![](_page_21_Figure_3.jpeg)

Figure 13: Weighted least squares adjusted linear trend model as a function of time (gray line) for the 9 estimates of the population mean by the sample mean ( $\overline{y}_s$ ) (black dots). The results for 10<sup>4</sup> samples. See the text for details.

#### <sup>512</sup> 5.3.2. Superpopulation

In this section, we document the bias of the sample mean  $\overline{y}_s$  in the context of both the preferential sampling process modeled by CPS and the process of spatial variation in abundance. The  $\xi p$ -expectation of  $\overline{y}_s$  is written as follows (we recall that the notation indicates the two sources of stochasticity involved, with subscript  $\xi$  for the superpopulation model and subscript p for the sampling process):

$$E_{\xi p}\left(\overline{y}_{s}\right) = E_{\xi}\left(E_{p}\left(\overline{y}_{s}\right)\right) \tag{18}$$

In our model, the population mean  $\overline{y}_{\mathcal{U}}$  is constant because we always allocate a fixed number of Mindividuals among a fixed number of N units (no demographic change). The  $\xi p$ -bias is therefore:

$$B_{\xi p}\left(\overline{y}_{s}\right) = E_{\xi p}\left(\overline{y}_{s}\right) - \overline{y}_{\mathcal{U}} \tag{19}$$

and the relative  $\xi p$ -bias is  $B_{\xi p}(\overline{y}_s)/\overline{y}_{\mathcal{U}}$ . The *p*-expectation of  $\overline{y}_s$  can be expressed by using the *p*-bias expression, which gives:

$$B_{\xi p}\left(\overline{y}_{s}\right) = E_{\xi}\left[B_{p}\left(\overline{y}_{s}\right) + \overline{y}_{\mathcal{U}}\right] - \overline{y}_{\mathcal{U}}$$

$$\tag{20}$$

If we are interested only in the value of the  $\xi p$ -bias (or the relative  $\xi p$ -bias), as mentioned in Section 4.4, we do not need to replicate the sampling process in our simulation since we know how to compute the *p*-bias analytically (Eq. 11 or Eq. B.11). However, it is necessary to simulate **y**-vector realizations to approximate the  $\xi$ -expectation that appears in the right term of expression (20).

In the following, we simulate  $10^4$  y-vectors for each  $\eta$ -value varying between  $\eta = 4$  (spatial gradient 526 in abundance from  $x_{\min}$  to  $x_{\max}$ ) and  $\eta = -4$  (gradient from  $x_{\max}$  to  $x_{\min}$ ) with a step of 0.25. We can 527 also compute the correlation in the model — assimilated here for simplicity to the  $\xi$ -expectation of the 528 population correlation — for each simulated situation. On average, under the model, the population 520 correlation varies between approximately 0.8 for  $\eta = 4$  at time t = 1 and -0.8 for  $\eta = -4$  at time t = 9, 530 with  $\rho_{\pi y} = 0$  for  $\eta = 0$  at time t = 5 (Fig. 14.a). The relative  $\xi p$ -bias ranges from approximately 531 -52.7% for  $\eta = -4$  at time t = 9 to 52.7% for  $\eta = 4$  at time t = 1 (Fig. 14.b). As a result, 532 a preferential sample set up at the beginning of the monitoring program — as explained above — 533 results in a downward trend (approximately linear) on average under the model, while the number of 534 individuals did not actually change over the  $\Delta$  period, but the spatial gradient was reversed. 535

![](_page_22_Figure_2.jpeg)

Figure 14: Approximation by Monte Carlo simulation ( $10^4$  realizations of the model) of the relation with time (t) of (a) the  $\xi$ -expectation of the population correlation between the propensity and the variable of interest and (b) the relative  $\xi p$ -bias of the sample mean.

### 536 6. Discussion

To date, the issue of nonprobability preferential sampling (as defined in Section 2.2) has not received sufficient attention in ecology. This may mean that most ecologists do not have a clear idea of the implications of this type of nonprobability sampling for their studies and monitoring programs. This observation led us to examine the issue in detail from a methodological perspective, through the example of a program aimed at estimating the status or trend of a finite population parameter.

First, we formalized the basic instances of without-replacement sampling processes by modeling them in terms of probability sampling designs. While we have applied this approach to spatial sampling, its versatility extends to temporal sampling, where the issue of preferential sampling may also arise (for example, migratory bird surveys are typically conducted at specific times of the year when bird migration is expected to peak).

Next, we used Monte Carlo simulations to study the bias of the sample mean as an estimator of 547 the population mean, first in the case of status assessment and second in the case of temporal trend 548 assessment. In the case of estimating the status of a population parameter (e.g., mean species richness), 549 we illustrated the risk of preferential sampling in terms of bias in the estimation of the population 550 mean since the relative bias may be unacceptable (e.g., greater than 10%), thereby leading to erroneous 551 conclusions. In addition, we showed that the estimation of the sampling variance was also biased when 552 simply using the formula for the variance of the mean  $s_y^2/n$ . We also noticed that the fact that the 553 sampling process results in fixed or variable sample sizes did not appreciably alter the results obtained 554 and is therefore not in itself a cause for concern. Thus, treating the sample size as fixed and estimating 555 parameters conditional on its value is a legitimate practice, at least for the question with which we 556 are concerned in this article, for a sufficiently large population size and moderate sampling fraction. 557 In the case of estimating the temporal trend of a population parameter such as mean abundance, the 558 simulated example shows that using preferential sampling at the beginning of the program to establish 559 permanent plots or sites can lead to the identification of a downward demographic trend when there 560 is a directional change in the spatial distribution of individuals, even though abundance is actually 561 constant over the monitoring period. 562

Hence, we clearly demonstrated that, without knowledge of the inclusion probabilities, the sampler 563 lacks awareness of the magnitude of the potential biases in either the sample mean or its sampling 564 variance. The simulated case for the trend assessment is a clear example of time-varying spatial 565 bias (Johnston et al., 2023, Sec. 3.1). The type of erroneous conclusions reached in this case has 566 already been clearly stated by Fournier et al. (2019) for abundance and Palmer (1993) for community 567 heterogeneity, but these authors referred to the phenomenon of *regression to the mean*, which differs 568 from what we have considered in this article. In a different ecological scenario than the one we 569 considered, McClure and Rolek (2023) noted that preferential sampling may delay the detection of 570 a downward trend. For biological conservation purposes, the common conclusion is that preferential 571 sampling can have a profound impact on trend estimation, either by highlighting a downward trend 572 where none exists (Type I statistical error) or by failing to detect (in time) a downward trend that 573 actually exists (Type II statistical error, insufficient statistical power). These two possibilities should be 574 considered when evaluating the statistical power of a nonprobability preferential sampling monitoring 575 program designed to detect downward trends. 576

In this article, a new perspective on biased site selection was introduced to quantitative ecology. 577 We have modeled preferential sampling processes using probability sampling designs for the purpose 578 of Monte Carlo simulation and also analytical formulation of the estimation bias. For status and trend 579 assessment, we caution against the use of preferential sampling in a nonprobability framework. We 580 showed that the bias in estimating the mean increases with the covariance between the propensity 581 and the variable of interest (species richness, abundance, etc.) and decreases with increasing sampling 582 effort. This is a simple but fundamental result that deserves greater recognition. In the following 583 sections, we take a closer look at the discussion points that emerge from our article. 584

#### 585 6.1. The scope of the simulated examples

The simulated examples in our article are intended to make our findings as clear as possible. Some objections can be raised: (1) real ecological studies and programs are not based on preferential sampling, so the problem documented in this article does not occur in practice; (2) the simulated situations bear no relation to reality.

Regarding point (1), it is clear that many programs actually use preferential sampling, at least 590 implicitly. Observers have a general tendency to select sites that are species rich (with rare species) 591 and/or high in abundance rather than opting for sites that may initially seem less interesting for 592 naturalistic observations. This behavior applies to both flora and fauna data collection (see, e.g., 593 Chytrý, 2001; Lepš and Śmilauer, 2007; Diekmann et al., 2007; Conn et al., 2017; Bowler et al., 2022). 594 Many examples can be provided from the ornithological world, given birds are among the best studied 595 of all animal groups. In the case of Galliforms, for example, a comprehensive review of data sources 596 on a global scale shows that, with the exception of atlases, data tend to be collected preferentially 597 from sites visited by tourists and bird specialists (biodiversity hotspots), while areas with few rare 598 species or protected areas are neglected (Boakes et al., 2010, p. 5). This results in a spatial coverage 599 problem typical of preferential sampling but which also involves convenience sampling because of access 600 difficulties (for logistical or political reasons). Another example is the Dutch Breeding Bird Monitoring 601 *Program* (BMP), which again relies on both convenience and preferential sampling since the observers 602 are free to choose their study areas, and in each habitat, they may prefer the most attractive sites, 603 i.e., those that are species-rich and have high bird densities (van Turnhout et al., 2008). An example 604 of a program that relies on both purposive and preferential sampling is the International Waterbird 605 Count (IWC) — a site-based counting scheme for monitoring waterbird numbers organized at the 606 supranational level by Wetlands International — where sites are defined by the judgment of national 607 coordinators and local observers and where decisions about which sites to count are based on their 608 relative importance (Delany, 2010, Sec. 4). These few examples illustrate that programs based on 609 nonprobability sampling of count sites do not fall under a single sampling type but generally involve 610 some degree of preferential sampling. 611

With respect to point (2), when estimating population trends in the context of global warming, 612 range shifts have already occurred in many places for different taxa and can be expressed in terms 613 of latitude, longitude, elevation or depth (see, e.g., McCarty, 2001; Parmesan, 2006, 2019; Lenoir 614 and Svenning, 2013, 2015; Dahms and Killen, 2023 and references cited therein). While real-world 615 situations may not be as extreme as the complete reversal of a spatial gradient in abundance as 616 simulated in this paper, the range shift phenomenon is undeniably real; moreover, its frequency is 617 likely to increase with global warming, as temperature and drought (for terrestrial ecosystems) are 618 the most limiting abiotic factors for many species. For birds, for example, range shifts in wintering 619 areas have been well documented (e.g., Maclean et al., 2008, Lehikoinen and Sparks, 2010; Lehikoinen 620 et al., 2013). For some species, this phenomenon may also act in synergy with a change in predation 621

pressure (Pokallus and Pauli, 2015). Changes in trophic interactions may contribute to a change in the geographic distribution of individuals. For example, this may be the case for the Eider Duck (*Somateria mollissima*), for which there is a possible shift from open islands to forested islands. Since the islands monitored are mostly open islands, the result can be the observation of a spurious downward trend (Ekroos et al., 2012, Sec. 4.2). Due to the gregarious behavior of migratory birds (for example) and communication between individuals, it is also possible that such shifts occur quickly.

In addition to the examples simulated in this article, as we have previously explained in detail in another context (Aubry et al., 2020, Sec. 5.1), the estimation by the sample mean is biased if the propensities vary significantly and are correlated with the variable of interest (e.g., species richness, abundance). This aligns well with the situation of preferential sampling described in this article. We note that a similar result occurs when estimating the variance of the sampled population, which may be a goal in itself (see Courbois and Urquhart, 2004), a topic not covered in this article.

#### 634 6.2. The bias of the sample mean

Beyond the results obtained by Monte Carlo simulation, we formally showed that the bias of the 635 sample mean as an estimator of the population mean can be written essentially as the population 636 covariance between the propensities and y-values divided by the mean propensity, i.e., the sampling 637 fraction in the case of a fixed-size sampling process or its expectation in the case of a variable-size 638 sampling process (Appendix B). Although Aubry et al. (2020) and Boyd et al. (2023) have recently 639 mentioned the key role played by the correlation between sample membership and the variable of 640 interest, to our knowledge, the analytical expression for the bias of the sample mean has so far remained 641 unknown to the ecological audience. In this respect, our paper fills a statistical gap in the field of 642 quantitative ecology. 643

The analytical expression for the bias (or similarly, the relative bias) of the sample mean can also 644 be found in the statistical literature specializing in the treatment of nonresponse (see Kalton and 645 Maligalig, 1991, Eq. 1.3; Särndal et al., 1992, p. 577, Eq. 15.6.3, 15.6. 4; Bethlehem, 1988, Eq. 646 3.5; Bethlehem, 1999, p. 129; Bethlehem, 2002, p. 276; Särndal and Lundström, 2005, p. 92; Brick 647 and Montaquila, 2009, Eq. 4; Bethlehem, 2009, p. 222; Bethlehem et al., 2011, p. 44; Haziza and 648 Beaumont, 2017, Eq. 3.4). An algebraically equivalent formula, but of less pedagogical interest, is 649 given by Groves et al. (2004b, Appendix) (see also Groves et al., 2004a, p. 182). We would like to 650 draw the reader's attention to the fact that this specialized literature is a methodological resource of 651 utmost interest that should be taken into account in quantitative ecology, as was also recently noted 652 by Chadwick et al. (2024, Box 1). 653

To summarize, when propensities vary among the sampling units, for a given spatial population and 654 variable of interest, there are three possible cases where using the sample mean may or may not result 655 in a biased estimation of the population mean. We illustrate these three cases with the help of Fig. 656 15, where the variances of the propensities and y-values do not change between the three situations 657 examined, with the fixed population being that of Fig. 12.1a. If there is a positive correlation between 658 the propensities and y-values (Fig. 15.1a), then the sampling is preferential, and the sample mean bias 659 is positive (overestimation) (Fig. 15.1b). If the correlation is zero, then, although the propensities 660 are variable (Fig. 15.2a), the sampling is nonpreferential, and there is no bias (Fig. 15.2b). If the 661 correlation is negative (Fig. 15.3a), then the sampling is antipreferential, and the bias is negative 662 (underestimation) (Fig. 15.3b). 663

![](_page_25_Figure_0.jpeg)

Figure 15: For the fixed population shown in Figure 12.1a, the bias of the sample mean according to the correlation between the propensities and y-values  $(R_{\pi y})$ . (1)  $R_{\pi y} = 0.8$ . (2)  $R_{\pi y} = 0$ . (3)  $R_{\pi y} = -0.8$ . (a) Scatter plot of the N = 2500 sampling units according to their propensity  $\pi$  and y-value. (b) Top panel: Example of a sample drawn by conditional Poisson sampling with the corresponding  $\pi$ -vector. Bottom panel: Relative bias of the sample mean (%) for the corresponding sampling process.

### 664 6.3. Reasons for using preferential sampling

Given the risk that preferential sampling poses to the validity of the conclusions drawn from the estimates, we need to ask why it is used. To answer this question, we must first distinguish between two situations, depending on whether the sampler (i) has a sampling frame or (ii) does not.

A sampling frame implies that the size of the spatial population is known. Therefore, it is possible, 668 at least in principle, to use a probability sampling design that allows for the knowledge of inclusion 669 probabilities and their incorporation into the estimators. The sampler can then estimate population 670 parameters within a *design-based* framework that is as objective as possible since there are no assump-671 tions about either the statistical distribution of the variable of interest or its possible spatial structure 672 (spatial trend and autocorrelation), unlike a *model-based* approach (see, e.g., de Gruijter and ter Braak, 673 1990; Brus and de Gruijter, 1993, 1997; Gregoire, 1998; Dumelle et al., 2022; Aubry and Francesiaz, 674 2022). In doing so, the sampler is aware of the statistical properties of the chosen estimators. This is 675 a fundamental difference from the concrete situation discussed in this article (see Section 2.4). 676

There may be several reasons for using preferential sampling within the design-based framework. Usually, the goal is to optimize the efficiency of the sampling strategy — in the sense of a pair formed by a sampling design and an estimator (e.g., Hedayat and Sinha, 1991, p. 24) — by minimizing the *p*-variance (sampling variance). For a fixed-size sampling design, an examination of the Sen-Yates-Grundy variance formula (e.g., Hedayat and Sinha, 1991, p. 48) — or what leads to the same

conclusion, that of its estimator, used in Eq. (6) — shows that the *p*-variance of the weighted estimator 682 is zero in the case of exact proportionality  $(\pi_i \propto y_i \text{ for } i \in \mathcal{U})$ . In practice, we do not know the y-values, 683 but we may know a size variable z which is positively correlated with y. Thus, we approach the ideal 684 proportionality relation by computing the inclusion probabilities as in Eq. (8). For a variable-size 685 sampling design (i.e., Poisson sampling), a similar result holds (see Särndal et al., 1992, pp. 86-87). 686 In a multispecies (i.e., multivariable) context, another reason may be to maximize the probability of 687 covering the spatial distributions of several species, as recently illustrated by Aubry et al. (2023). The 688 goal is to optimize the sampling effort to collect data for a maximum number of species in a limited 680 number of site visits while remaining within the framework of design-based estimation. 690

If the sampler lacks a sampling frame, the selection of the sample of spatial units is typically not 691 randomized — at least as we consider the sampling randomization in this article. Even leaving aside 692 the question of randomization, specifying in advance the spatial units to be visited in such programs 693 can be challenging. In these cases, the preferential nature of the sampling process is then no longer 694 dictated by statistical considerations but rather by concerns about the adherence of the observers — 695 often volunteers — to the program. The preferential nature of the sampling process often results from 696 the collective action of observers who tend to select units in a similar manner (see, e.g., ter Steege 697 et al., 2011). 698

In contrast to preferential sampling, antipreferential sampling may refer to preferential inclusion in the sample of species-poorest units, the units with the lowest abundances or occupancy probabilities. The objectives may be to overrepresent the distribution margins of a species in the sample, to study the effect of buffer zones in the establishment of protected areas or to study the natural recovery of degraded ecosystems.

#### <sup>704</sup> 6.4. Reasons for avoiding preferential sampling in trend assessment

In general, the use of (i) a permanent sample selected by preferential sampling and (ii) the sample 705 mean as an estimator of the population mean is a sampling strategy that should not be followed for 706 trend assessment. We have illustrated the problem of such a strategy that leads to the conclusion 707 that there is a downward trend even when there is no trend at all. This spurious trend is caused 708 by the concomitance of preferential sampling at the beginning of the program and a change in the 709 geographic distribution of individuals during the monitoring period. In this case, if the propensities 710 (inclusion probabilities) had been known — as would have been the case if the fixed-size sample 711 had been obtained by applying an unequal-probability sampling design such as *conditional Poisson* 712 sampling (for example) — then it would have been possible to estimate the mean abundance without 713 bias, as well as the variance of the estimator (see Section 3.5). This would have attenuated the 714 problem of spurious downward trend detection. However, in the simulated case, the sampling variance 715 increases over time as the correlation between the propensities and the y-values decreases, vanishes 716 (null correlation) and then changes sign (negative correlation) (Fig. 14.a). Thus, even if the inclusion 717 probabilities were known and incorporated into the estimators, the strategy used was inappropriate 718 and risky. 719

When assessing a temporal trend in abundance (for example), using a permanent sample of units 720 selected by preferential sampling at the beginning of the monitoring program can potentially lead to 721 three types of erroneous conclusions: (i) highlighting a downward trend when there is no demographic 722 change (the case illustrated in this article); (ii) attenuating an upward trend (e.g., when there is a 723 colonization of new habitat patches that are almost unrepresented in the sample, without changes in 724 habitat patches previously occupied, which constitute the bulk of the sample); and (iii) exaggerating a 725 downward trend (e.g., in the case of a density-dependent decline, more pronounced in units where there 726 were many individuals initially, which represents the bulk of the sample). In the case of antipreferential 727 sampling, the results are reversed. The only case for which this sampling strategy does not lead to a 728 time-varying spatial bias is when the change in abundance is uniform across the spatial domain under 729 consideration, i.e., when it occurs at the same rate for all sampling units. It is understandable that 730 this case is the exception rather than the rule in ecology. 731

## 732 6.5. Consideration of the sampling process in ecology

Like others (e.g., Elzinga et al., 2001, p. 116; Albert et al., 2010; Smith et al., 2017; Aubry et al., 2020; Boyd et al., 2023), we recommend the use of probability sampling whenever possible to control for inclusion probabilities. This recommendation applies whether the planned inference is design-based or model-based. It is often thought that the use of a probability sampling design is not applicable in ecology, especially on a large scale. While we cannot deny that using probability sampling may rate design indeed be difficult or even unfeasible (see, e.g., Roleček et al., 2007), there are counterexamples that illustrate that this approach may be possible, at least for certain taxonomic groups and environments
(e.g., Pavlacky et al., 2017; van Wilgenburg et al., 2020; Aubry et al., 2023).

When it is not possible to control for inclusion probabilities, it is important to be aware of the 741 source of bias that preferential sampling may represent and to discuss it when communicating the 742 results of a study or program. We agree with Boyd et al. (2023) on this point while recognizing that 743 this is the least one can do. A further step forward is the recognition that statistically sound use of 744 the data obtained by preferential sampling necessarily requires consideration of the sampling process, 745 whether the chosen inferential framework is frequentist or Bayesian. This topic is an active area of 746 research in survey sampling and spatial or ecological statistics, which goes beyond status and trend 747 assessment (see Journel, 1994; Diggle et al., 2010; Pati et al., 2011; Zidek et al., 2014; Pan et al., 2015; 748 Grisotto et al., 2016; Cecconi et al., 2016; Conn et al., 2017; Irvine et al., 2018; Watson et al., 2019; 749 Dinsdale and Salibian-Barrera, 2019; Pennino et al., 2019; da Silva Ferreira, 2020; Olea, 2021; Gray 750 and Evangelou, 2023; Vedensky et al., 2023). Such research is critical for existing ecological programs 751 that were not built on the basis of probability sampling, but addressing the statistical approaches that 752 can be considered a *posteriori* to deal with the problem of bias induced by preferential sampling is 753 beyond the scope of this article. Nevertheless, for a recent introduction of this topic in ecology, we 754 refer the reader to Boyd et al. (2024). 755

Regarding sampling processes in ecology, the main message of this article can be summarized as 756 follows: "If we do not know the sampling process, we have no idea what we are doing with the data (at 757 least from a statistical perspective)". While this statement may seem self-evident to statisticians, an 758 examination of the ecological literature reveals a disturbing reality: the somewhat widespread neglect 759 of this basic premise. In practice, statistical analysis and modeling are often performed without due 760 consideration of the sampling process. This is not a problem if one is interested only in the data 761 from the sample at hand, but it is a guarantee of poor statistical inference when the results must be 762 extended beyond the sample. We believe that the oversight about the sampling process at work is a 763 major contributor to the misuse of statistical methods in ecology (analytical crisis, Chadwick et al., 764 2024). 765

#### 766 6.6. Focusing on the terminology

Scientific discourse requires accurate, monosemous and, ideally, stable and shared terminology (see, 767 e.g., Schuster, 2020, Sec. 2). Unfortunately, there is no consensus on the terminology used to describe 768 preferential sampling or related concepts. Since the lack of common understanding and naming of 769 concepts is a major barrier to communication among researchers, we felt it important to devote part 770 of the discussion in this article to this issue, to ensure proper and operational transfer in quantitative 771 ecology. We fully agree with Hall et al. (1997) that "[ $\cdots$ ] if we want to advance [ $\cdots$ ] ecology, we must be 772 sure that the fundamental concepts with which we work are well defined, and hence, well understood". 773 The term *preferential sampling* has been used in the literature to denote spatial sampling when 774 it is neither regularly nor randomly distributed across the study area (e.g., Goovaerts, 1997, Sec. 775 4.1.1), as a synonym for *purposive sampling* of typical units (e.g., Orlóci, 1975, pp. 10, 12; Podani, 776 1984; Roleček et al., 2007; Swacha et al., 2017), for convenience sampling, mainly for ease of access 777 (e.g., Clifford et al., 2011; Mentges et al., 2021). Other distinct meanings can also be found in the 778 literature. For example, Merckx et al. (2011) use the term preferential sampling to refer to visiting 779 some sites more frequently than others. In numerous articles, the term is not even defined by the 780 authors who use it, which is problematic because it can mean different things to different people. 781 Moreover, it covers different aspects that need to be distinguished because they may have different 782 statistical consequences, requiring different statistical approaches to be properly handled. In this 783 article, we use the term *preferential sampling* in a precise statistical sense to denote the existence of 784 a nonnegligible correlation between the propensities of the units to be sampled and the values of the 785 variable of interest. It is used in a similar way as McClure and Rolek (2023), except that our definition 786 is more deeply rooted in statistics. 787

Preferential (or antipreferential) sampling is a special case of what is more generally known in the 788 literature as *biased selection* in the sense that the sampling process produces samples in which some 789 parts of the population are underrepresented while others are overrepresented (Zarkovich, 1966, p. 75). 790 The term *selection bias* may be used by some authors to refer to the same idea (e.g., Eklund, 1959, Sec. 791 3.2), in the sense of samples or data biased by selection (sample selection bias or selection-biased data). 792 This terminology refers to one of the usual meanings given to the term *bias*, i.e., that associated with a 793 distortion or deformation (i.e., of a study, a result, a conclusion, etc.). However, the expression *selection* 794 bias can also refer to the technical statistical meaning of the term bias when it concerns an estimator 795 (e.g., Kotz et al., 2006, p. 483). We note that biased selection does not necessarily lead to biased 796 estimation (e.g., Overton and Stehman, 1995, Example 6; Aubry et al., 2020, Sec. 5.1), for example, 797

when using a weighted estimator based on inclusion probabilities (e.g., using the Horvitz-Thompson estimator). Conversely, the absence of biased selection does not guarantee unbiased estimation (e.g.,
Stuart, 1984, Sec. 6). This terminology can therefore be confusing; the most important thing is to know what one is talking about and to use the terms consistently.

The term *site-selection bias* (e.g., Irvine et al., 2018; Fournier et al., 2019; Mentges et al., 2021) is another synonym for *preferential sampling* in the meaning used in this article, or includes it as a subcase (McClure and Rolek, 2023). The two synonymous terms *response biased sampling* or *response selective sampling* can also be found in the literature to refer to (or include) preferential sampling (e.g., Lawless, 1997; Lawless et al., 1999; Scott and Wild, 2011). However, these two terms seem to be used almost exclusively by (some) statisticians.

By using statistical terminology from the field of missing data (see, e.g., Allison, 2002; Enders, 2010; Molenberghs et al., 2015; Little and Rubin, 2019), whenever sampling is preferential (Fig. 15.1) or antipreferential (Fig. 15.3), the missingness mechanism is said to be *nonignorable*, and the data are said to be *missing not at random* (MNAR; also referred to as *not missing at random* or NMAR, see Little and Rubin, 2019, p. 28, Note 1). If the propensities vary minimally or vary significantly but are not correlated with the variable of interest (Fig. 15.2), then the mechanism is said to be *ignorable*, and the data are said to be *missing completely at random* (MCAR).

Boyd et al. (2023, 2024) use the term *representative* to refer to nonpreferential sampling, i.e., 815 when sample membership is uncorrelated with the variable of interest. Given the already widely 816 polysemous nature of the terms representative sampling and representative sample (see, e.g., Kruskal 817 and Mosteller, 2006; Bethlehem, 2009, Sec. 2.4.1), this new definition is likely to create more confusion 818 than clarification. Since sampling is a mechanism that produces missing data, it is more appropriate 819 to use the terminology used in that field and simply refer to sampling as *ignorable* or *nonignorable*. 820 as the case may be. When dealing with statistical issues in ecology, it is appropriate for clarity and 821 consistency to refer to the vocabulary used in statistics, as Irvine et al. (2018) do, for example. 822

#### **7.** Perspectives

The points discussed in this article relate to spatial sampling in ecology — that is, sampling of 824 spatial units to study or monitor ecological phenomena — but they are quite general from a statistical 825 perspective and concern broader topics than those covered here (e.g., Aubry et al., 2020). We have used 826 the (seemingly simple) example of status and trend assessment, but the issue of spatial preferential 827 sampling has also been highlighted in other fields (e.g., spatial prediction, Gelfand et al., 2012). We 828 have considered preferential spatial sampling — which is undoubtedly the most obvious instance of 829 preferential sampling in ecology — but temporal sampling should also be accounted for in practice. 830 For example, if the sampling period targets an annual peak in abundance but the phenology of the 831 species of interest is gradually changing with global warming, then there is a serious risk of introducing 832 a time-varying temporal bias, which is the counterpart of the time-varying spatial bias illustrated in 833 this article. 834

We believe that the analytical formula governing the bias of the sample mean (Eq. 11) is fundamental and should be familiar to ecologists and wildlife biologists. Despite the recognized importance of sampling, it is paradoxical that it receives so little attention overall in quantitative ecology, a field largely dominated by modeling. Even in the case of a model-based approach, sampling issues remain central, as they largely determine the ability of the model fitted to the sample data to reliably estimate or predict the quantities of interest. Further attention and work are needed in this area, as we believe this topic is critical to the credibility of the results published in the ecological literature.

Demonstrating the risk of the sampling strategy documented in this article for trend assessment 842 is undoubtedly useful, but it is even more useful to suggest strategies that are as robust as possible 843 for detecting a trend that is not spurious, not attenuated or not exaggerated to guide the readers in 844 their choices when designing a monitoring program. It is not simply a matter of drawing inspiration 845 from existing programs but of justifying strategies in light of ecological, statistical and operational 846 perspectives. Quoting Greenwood (2003): "In designing surveys, however, we strike the balance be-847 tween theoretical robustness and practicality: just as there is no point in running a survey so biased 848 that the data are uninterpretable, there is no point in designing one that is so theoretically perfect that 849 it is impossible to conduct.". This may be the subject of future articles. 850

Another perspective concerns the situation in which probability sampling is ruled out (for various reasons). A statistical approach for dealing with nonprobability sampling data is to use a model, be it frequentist or Bayesian. This approach is known as the *model-based* approach. Strictly speaking, the *design-based* approach cannot be used, precisely because the sampling was not conducted by using a probability sampling design. However, a remaining question arises: Just as we have used probability

sampling designs to model sampling processes for simulation purposes, to what extent can they also 856 be used for statistical inference purposes? This approach differs from design-based inference in the 857 strictest sense of the term because, in this case, the inference is actually based on modeling the sampling 858 process using a probability sampling design. This approach may be called *quasi-randomization* in the 859 sense given by Oh and Scheuren (1983) or pseudo design-based (Baker et al., 2013). This rather 860 unusual topic in quantitative ecology (see Boyd et al., 2024 for a primer) should be explored in detail 861 in future articles and echoes the question posed by Boyd et al. (2023)" What other methods for making 862 inferences from nonprobability samples exist, and how reliable are they with real data?". Such a study 863 can be undertaken concurrently with an examination of the robustness of predictions based on a 864 superpopulation model, a topic recently illustrated by Aubry and Francesiaz (2022). 865

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### 870 Appendix A. Notational conventions

We denote the mean of a variable x on a finite set A:

$$\overline{x}_A = \frac{1}{|A|} \sum_{i \in A} x_i$$

where |A| is the size (cardinality) of A.

We denote the *adjusted variance* of x on the finite population  $\mathcal{U}$ :

$$S_x^2 = \frac{1}{N-1} \sum_{i \in \mathcal{U}} (x_i - \overline{x}_{\mathcal{U}})^2$$

With this notation, the population variance has N - 1 as the denominator, following the convention used in survey sampling theory (see Cochran, 1977, p. 23). When the finite population is viewed as randomly drawn from a superpopulation, using N - 1 as the denominator also makes sense since the population variance is then an unbiased estimator of the superpopulation variance (e.g., O'Neill, 2014, p. 283). Similarly, the (adjusted) sample variance has the denominator  $n_s - 1$ :

$$s_x^2 = \frac{1}{n_s - 1} \sum_{i \in s} \left( x_i - \overline{x}_s \right)^2$$

where  $n_s$  is the sample size, which may or may not depend on the realized sample s (variable or fixed sample size).

We denote the (adjusted) covariance between two variables x and y on the finite population  $\mathcal{U}$ :

$$S_{xy} = \frac{1}{N-1} \sum_{i \in \mathcal{U}} \left( x_i - \overline{x}_{\mathcal{U}} \right) \left( y_i - \overline{y}_{\mathcal{U}} \right)$$

The correlation between two variables x and y on the finite population  $\mathcal{U}$  is defined as  $R_{xy} = \frac{S_{xy}}{S_{xy}} \frac{[S_x S_y]}{[S_x S_y]}$ .

In a superpopulation model denoted  $\xi$ , the  $\xi$ -variance of a variable x is noted  $\sigma_x^2$ , and the  $\xi$ covariance between two variables x and y is denoted  $\sigma_{xy}$ . The  $\xi$ -correlation between x and y is then defined as  $\rho_{xy} = \sigma_{xy} / [\sigma_x \sigma_y]$ . This is the definition used in Section 4. Mathematically,  $R_{xy}$  is not an unbiased estimator of  $\rho_{xy}$ ; that is,  $\rho_{xy} \neq E_{\xi}(R_{xy})$ . However, the bias is negligible in the case of a sufficiently large population. In Section 5, for convenience, we use the approximation  $\rho_{xy} \approx E_{\xi}(R_{xy})$ .

### <sup>889</sup> Appendix B. Bias of the sample mean

Let a finite population  $\mathcal{U}$  of size N be sampled by a without-replacement sampling design with inclusion probabilities  $0 < \pi_k \leq 1$  ( $k \in \mathcal{U}$ ). Let s be a sample of size  $n_s$  drawn from  $\mathcal{U}$ . Introducing the indicator variable  $I_k = 1$  when unit  $k \in \mathcal{U}$  is included in sample s and  $I_k = 0$  otherwise, the sample mean is written:

$$\overline{y}_s = \frac{1}{n_s} \sum_{k \in s} y_k = \frac{1}{n_s} \sum_{k \in \mathcal{U}} I_k y_k = \frac{\sum_{k \in \mathcal{U}} I_k y_k}{\sum_{k \in \mathcal{U}} I_k}$$
(B.1)

#### Appendix B.1. Variable-size sampling design 894

In a variable-size sampling design, the sample size  $n_s$  is a random variable of expectation: 895

$$E_p(n_s) = E_p\left(\sum_{k \in \mathcal{U}} I_k\right) = \sum_{k \in \mathcal{U}} E_p(I_k) = \sum_{k \in \mathcal{U}} \pi_k$$
(B.2)

By using the exact expression for the expectation of a ratio of two random variables (see Midzuno, 896 1950; Koop, 1951, 1972), the expectation of the sample mean (B.1) can be written as follows: 897

$$E_{p}(\overline{y}_{s}) = E_{p}\left(\frac{\sum_{k \in \mathcal{U}} I_{k} y_{k}}{\sum_{k \in \mathcal{U}} I_{k}}\right) = \frac{E_{p}\left(\sum_{k \in \mathcal{U}} I_{k} y_{k}\right)}{E_{p}\left(\sum_{k \in \mathcal{U}} I_{k}\right)} + \epsilon$$
(B.3)

/

with 898

$$\epsilon = \mathbf{E}_p \left[ \left( \sum_{k \in \mathcal{U}} I_k y_k \right) \left\{ \left( \sum_{k \in \mathcal{U}} I_k \right)^{-1} - \left( \sum_{k \in \mathcal{U}} \pi_k \right)^{-1} \right\} \right]$$
(B.4)

The  $\epsilon$ -term is usually negligible, but this is not the case for very small populations. Therefore, from 899 Eq. (B.3), we can use the following approximation: 900

$$E_{p}(\bar{y}_{s}) \approx \frac{E_{p}\left(\sum_{k\in\mathcal{U}}I_{k}y_{k}\right)}{E_{p}\left(\sum_{k\in\mathcal{U}}I_{k}\right)} = \frac{\sum_{k\in\mathcal{U}}\pi_{k}y_{k}}{\sum_{k\in\mathcal{U}}\pi_{k}} = \frac{1}{N\overline{\pi}_{\mathcal{U}}}\sum_{k\in\mathcal{U}}\pi_{k}y_{k}$$
(B.5)

with an average inclusion probability of  $\overline{\pi}_{\mathcal{U}} = N^{-1} \sum_{k \in \mathcal{U}} \pi_k$ . Therefore, the bias is approximated as 901 follows: 902

$$B_{p}(\overline{y}_{s}) = E_{p}(\overline{y}_{s}) - \overline{y}_{\mathcal{U}} \approx \frac{1}{N\overline{\pi}_{\mathcal{U}}} \sum_{k \in \mathcal{U}} \pi_{k} y_{k} - \overline{y}_{\mathcal{U}} = \frac{1}{\overline{\pi}_{\mathcal{U}}} \left[ \frac{1}{N} \sum_{k \in \mathcal{U}} \pi_{k} y_{k} - \overline{\pi}_{\mathcal{U}} \,\overline{y}_{\mathcal{U}} \right] = \frac{1}{\overline{\pi}_{\mathcal{U}}} \operatorname{Cov}_{\pi y} \quad (B.6)$$

with  $\operatorname{Cov}_{\pi y} = (N-1)S_{\pi y}/N$  where  $S_{\pi y}$  is the adjusted covariance (Appendix A). Therefore, we can 903 also write: 904

$$B_p(\overline{y}_s) \approx \frac{Cov_{\pi y}}{\overline{\pi}_{\mathcal{U}}} = \frac{(N-1)}{N} \frac{R_{\pi y} S_{\pi} S_y}{\overline{\pi}_{\mathcal{U}}}$$
(B.7)

For  $(N-1)/N \approx 1$ , we obtain the approximate expression: 905

$$B_p(\overline{y}_s) \approx \frac{R_{\pi y} S_\pi S_y}{\overline{\pi}_{\mathcal{U}}}$$
(B.8)

#### Appendix B.2. Fixed-size sampling design 906

With a fixed-size sampling design, the sample size is the constant  $n = \sum_{k \in \mathcal{U}} \pi_k$ , and the sample 907 mean (B.1) can therefore be written as follows: 908

$$\overline{y}_s = \frac{\sum_{k \in \mathcal{U}} I_k y_k}{\sum_{k \in \mathcal{U}} \pi_k} \tag{B.9}$$

<sup>909</sup> The expectation of the sample mean is then:

$$E_{p}(\overline{y}_{s}) = \frac{E_{p}\left(\sum_{k\in\mathcal{U}}I_{k}y_{k}\right)}{\sum_{k\in\mathcal{U}}\pi_{k}} = \frac{\sum_{k\in\mathcal{U}}\pi_{k}y_{k}}{\sum_{k\in\mathcal{U}}\pi_{k}} = \frac{1}{N\overline{\pi}_{\mathcal{U}}}\sum_{k\in\mathcal{U}}\pi_{k}y_{k}$$
(B.10)

 $_{910}$   $\,$  The bias is therefore written exactly:

$$B_p(\overline{y}_s) = \frac{\operatorname{Cov}_{\pi y}}{\overline{\pi}_{\mathcal{U}}} = \frac{(N-1)}{N} \frac{R_{\pi y} S_{\pi} S_y}{\overline{\pi}_{\mathcal{U}}}$$
(B.11)

<sup>911</sup> For  $(N-1)/N \approx 1$ , we again obtain the approximate expression (B.8).

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